



More Than Meets the Eye: Unraveling the Interactions Between Skin Microbiota and Habitat in an Opportunistic Amphibian

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Received: 5 November 2024 / Accepted: 2 January 2025
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Abstract

With amphibians still holding the record as the most threatened class of terrestrial vertebrates, their skin microbiota has been shown to play a relevant role in their survival in a fast-changing world. Yet little is known about how abiotic factors associated with different aquatic habitats impact these skin microorganisms. Here we chose the yellow-bellied toad (*Bombina variegata*), a small anuran that colonizes a wide range of wetland habitats, to investigate how the diversity and composition of both its bacterial and fungal skin communities vary across different habitats and with water characteristics (temperature, pH, and dissolved oxygen) of these habitats. Skin microbiota was sampled from 14 sites in the Province of Trento (Italy), including natural pools, ephemeral ponds, irrigation tanks, and farm ponds. Interestingly, the diversity of the two microbial components was also highly correlated. Close associations between both the diversity and composition of water and skin communities were noted for each habitat and sampling site, suggesting that water bodies actively contribute to the skin microbiota assemblage. In addition, water pH, temperature, and dissolved oxygen affected both bacterial and fungal diversity of skin. We confirmed the presence of *Batrachochytrium dendrobatidis* in skin samples of animals collected from eight waterbodies, as well as more than 60 microbial taxa previously associated with resistance to this pathogen. We concluded that both skin bacterial and fungal communities appear to be influenced by each other as well as by environmental communities and conditions, and these relationships connecting the whole ecosystem should be considered in future research concerning amphibian conservation.

Keywords Amphibian · Microbiota · Mycobiota · Fungi · *Bombina* · *Batrachochytrium dendrobatidis*

Introduction

As amphibian populations continue to decline worldwide, mainly due to human-driven habitat modifications and emerging diseases [1], skin microbiota is of particular

interest as both a possible means of adaptation to the rapidly changing environment and a defense against pathogens [2–4]. A substantial body of amphibian skin microbiota research has been generated in the last 10 years, ranging from the characterization of bacterial taxa [e.g., 5] to the assessment of microbiota variation between species and life stages [6, 7] or through time [8, 9]. As several threatened species are involved in ex situ captive-breeding programs, the impact of captive rearing on skin microbiota has also been summarized in a recent review [10]. Unsurprisingly, several studies have focused on the interaction between amphibian skin microbiota and known pathogens, especially the deadly chytrid fungus *Batrachochytrium dendrobatidis* (Longcore, Pessier & Nichols, 1999; hereafter, *Bd*). This pathogen has led to population declines and local extinctions of several amphibian species and is considered one of the major threats to amphibian conservation [11]. Such studies have shown how bacterial skin communities

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can influence the resistance of wild amphibian individuals to *Bd* infection [e.g., 12] and conversely, the impact of fungal infection on skin bacterial composition and diversity [13]. Several skin-associated bacteria are known to produce antifungal metabolites, effectively preventing *Bd* infections [14]. Interestingly, the possibility of translocating these taxa from one amphibian species to another has recently been tested as a means of increasing amphibian resistance to *Bd* [15]. However, only very recently has amphibian microbiota research started characterizing fungal skin communities (mycobiota) and their potential role in disease resistance [16–18]. In addition, the complex interplay of both bacterial and fungal communities is still largely unexplored [19, 20].

Amphibian skin bacterial communities are known to be influenced by several factors, including the composition of the environmental microorganism communities, which constitute an important source of bacteria and fungi for maintaining skin communities [21, 22], and habitat quality [23]. On a global scale, Kueneman and colleagues [24] have shown that amphibian microbiota changes in relation to macrohabitat, annual climatic trends such as temperatures and precipitation, and elevation. However, at a finer scale, the relative importance of specific abiotic variables like water temperature and pH to the observed variations in skin microbiota composition has only very recently begun to be addressed, and never for mycobiota [3 and references therein; 25].

Given the above gaps in knowledge, this study aimed to (i) characterize the skin mycobiota of a widespread amphibian and investigate, in four freshwater habitat subtypes, the associations between (ii) bacterial and fungal components of skin, (iii) skin microbiota and water microbiota, and (iv) skin microbiota and abiotic variables, as well the incidence of *Bd*. We chose to focus on *Bombina variegata* (Linnaeus, 1758) due to its ability to colonize a wide range of natural and constructed wetland habitats, although preferred reproductive sites are exposed temporary pools at the edge of woodlands [26]. Despite being widely distributed across central and southern Europe and classified as Least Concern (LC) in the IUCN Red List [27], yellow-bellied toad populations are locally in decline. Such is the case in the chosen study site, the mountainous Province of Trento, Italy [28], probably due to habitat loss because of urbanization and wetland pollution [29]. Hence, the species is listed as Vulnerable (VU) in the Provincial Red List [28]. The skin microbiota (including mycobiota) has never been studied previously, but infections of *Bd* on this species have been reported over much of its distribution [e.g., 30, 31], although never for Italian populations [32], nor have mass mortality events nor evident signs of disease ever been observed in the study area [33].

Materials and Methods

Skin Swabs, Water Samples, and Parameter Collection

Animal handling procedures for research purposes were authorized by the Italian Ministry of the Environment according to DPR 357/97 on April 7th, 2021 (Prot. ISPRA 16759 of 6/4/21; Authorization no. 0035657 of 7/4/21 by the Ministry of the Environment). We sampled *B. variegata* individuals from 14 sites in the Province of Trento, including five of the nine known reproductive habitat subtypes of this species described for this area by Endrizzi et al. [28]. These freshwater habitat subtypes were chosen to represent the ecological diversity of the species while also taking into account the known abundance, as some sites support very small and elusive populations (cfr. Figure 13 in [28]). Therefore, sampled sites included two “fluvial habitat subtypes” (FLU), four “valley bottom ecotone habitat subtypes” (VBE), four “agricultural habitat subtypes” (AGR) and four “pasture habitat subtypes” (PAS). General descriptions of these sites can be found in Table 1, as well as their respective location, number of samples collected, and water parameters. The geographical distribution of the sampled sites is represented in Fig. 1.

Six to 15 apparently healthy adult individuals per site were sampled between April and July 2022 (Table 1). Field workers wore sterile gloves, changing them between each toad, which was captured by hand or using a dip-net, rinsed with sterile water to remove transient bacteria, and swabbed with a GenoTube Livestock Swab (Thermo Fisher Scientific, USA), gently stroking their ventral and axillary areas for 20 s. Two water samples (200–500 ml) were collected per site from the shore and filtered with Sterivex GP Filter units (pore size 0.22 μm , Millipore cat. no. SVGPL10RC, Merck KGaA, Darmstadt, Germany) following [34]. Water temperature and dissolved oxygen were measured from the shoreline with an oximeter model Seven2Go DO (Mettler Toledo), and pH with a pH meter model Seven2Go pH/Ion meter S8 (Mettler Toledo). Site elevation was measured using Google Earth Pro version 7.3.6.9345, 2022. All samples (146 swabs and 28 filters) were transported at ambient temperature [35] to the Animal, Environmental and Antique DNA Platform at the Fondazione E. Mach (FEM) within 1 h of sampling, and then stored at $-20\text{ }^{\circ}\text{C}$ until DNA extraction.

Laboratory Workflow

All laboratory procedures were carried out under BSL2 biological hoods at FEM. DNA extraction from skin swabs was performed using the DNeasy Blood & Tissue Kit (Qiagen, Germany) following the Quick-Start standard protocol,

Table 1 General description of the sites where *Bombina variegata* individuals were collected, with the habitat subtype and water parameters (altitude and water pH, dissolved oxygen, and temperature) as well as number of samples (individuals and water) collected per site

Habitat code ^a	Site ID ^b	Location	No. of skin samples ^c	No. of water samples ^d	Altitude a.s.l. (m)	Water pH	Dissolved oxygen (mg/L)	Water temp. (°C)
FLU	AMB-1_1	Pozzolago, Lona Lases	10	2	411	7.76	7.86	18.6
FLU	AMB-1_2	Falesia di Prà, Segonzano	9	2	499	8.73	8.73	20.9
VBE	AMB-2_1	San Michele all'Adige	10	2	210	7.52	0.44	13.2
VBE	AMB-2_2	Maso delle Part- Maso Inon Mezzolombardo	15	2	211	7.15	5.78	18.1
VBE	AMB-2_3	Zambana	6	2	199	8.29	7.88	31.1
VBE	AMB-2_4	Mezzolombardo	10	2	209	7.23	1.94	19.8
AGR	AMB-3_1	Verla 1, Giovo	10	2	471	8.45	8.1	22
AGR	AMB-3_2	Verla 2, Giovo	11	2	470	9.15	10.61	22.7
AGR	AMB-3_3	Cembra 1 & 3, Lisignago	10	2	486	8.06	5.78	25.7
AGR	AMB-3_4	Lisignago 1	11	2	499	6.83	0.05	22.8
PAS	AMB-4_2	Monte Baldo near Postemon, Brentonico	11	2	1466	7.32	2.42	23.5
PAS	AMB-4_3	Postemon, Monte Baldo, Brentonico	12	2	1379	7.35	3.28	19.4
PAS	AMB-4_4	Malga Campe, Monte Baldo, Brentonico	11	2	1306	7.21	1.01	19.2
PAS	AMB-4_5	Malga Palazzo, Besenello	11	2	1533	6.96	13.72	29.1

^aFLU: two natural pools bordering the Avisio River in the Adige Valley (i.e., seasonal accumulations of water which form in the narrow river floodplains during periods of high flow, typically in spring and summer, “pozze laterali fluviali” in [28]), VBE: four ephemeral ponds (puddles forming in wheel ruts, “raccolte d’acqua temporanee” in [28]) and artificial drainage ditches (“fossi agricoli”, [28]) in apple orchards in the Piana Rotaliana, AGR: irrigation tanks in the vineyards in Val di Cembra (concrete structures for water collection, with a perimeter of 2–6 m, “vasche agricole” in [28]), and PAS: four high elevation farm ponds on Monte Baldo, Monte Bondone and the Altopiano della Vigolana (basins for watering grazing livestock, “pozze d’alpeggio” in [28])

^bIdentification codes for each sampling site

^cNumber of *B. variegata* skin samples collected (one per individual) for microbiota analysis

^dNumber of water samples collected for water microbiota analysis. Water pH, dissolved oxygen, and water temperature were measured near the shoreline, where these animals were found

including elution of extracted DNA in 200 µl Buffer AE. The lysis step was carried out overnight to maximize extraction of bacteria and fungi. Filter samples were extracted with the DNeasy PowerWater Sterivex Kit (Qiagen, Germany), following the manufacturer’s instructions (with a final elution in 100 µl of Solution EB). Each extraction batch included one negative control (sterile water only). Extracted DNA was stored at –20 °C until amplification.

The bacterial V3-V4 fragment of 16S rRNA gene was amplified with primers 341F (CCTACGGGNGGCWGC AG) and 805R (GACTACNVGGGTWTCTAATCC) [36, 37] (expected size ~ 460 bp; [38]) anchored with Illumina forward and reverse overhang adapters (<https://support.illumina.com/documents/documentation/>). The reaction mixture consisted of 12.5 µl of 2 × KAPA HiFi HS ReadyMix, forward and reverse primer to a final concentration of 0.32 µM, 3 µl of template DNA, and sterile H₂O to a final volume of 25 µl. The PCR program followed the KAPA Kit protocol (31 cycles and annealing temperature of 55 °C). For fungal metataxonomy, the ITS1 region was amplified with primers ITS5 (GGAAGTAAA AGTCGTAACAAGG) and ITS2 (GCTGCGTTCTTCATC

GATGC) [39] as described above. The expected size of the ITS1 fragment is usually around 250–400 bp (e.g., [40–42]). PCR reactions were carried out in a reaction mixture consisting of 10 µl of Promega Flexi Buffer 5X, forward and reverse primer to a final concentration of 0.2 µM, 1 µl of dNTP’s 10 mM each, 0.25 µl of Promega-GoTaq HS G2 5U/µl, 5 µl of MgCl₂ 25 mM, 5 µl of template DNA and sterile H₂O up to a volume of 50 µl. The PCR program followed the Promega Taq Kit parameters (40 cycles and annealing temperature of 55 °C). Each round of PCR included three extraction blanks and two to three PCR blanks (reaction mixture only). All amplification products were then purified with the CleanNGS Kit CNGS-0050 (CleanNA, The Netherlands) following the manufacturer’s instructions, and sequenced at the FEM Sequencing and Genotyping Platform with paired-end sequencing (2 × 300 base pairs, bp) on an Illumina Miseq (Illumina, San Diego, CA). The Illumina sequencing protocol therefore covered both target regions, based on our expected amplicon sizes. The sequencing depth per sample was 100,000 reads per sample for bacteria and 30,000 for fungi (following in part [41, 42]).

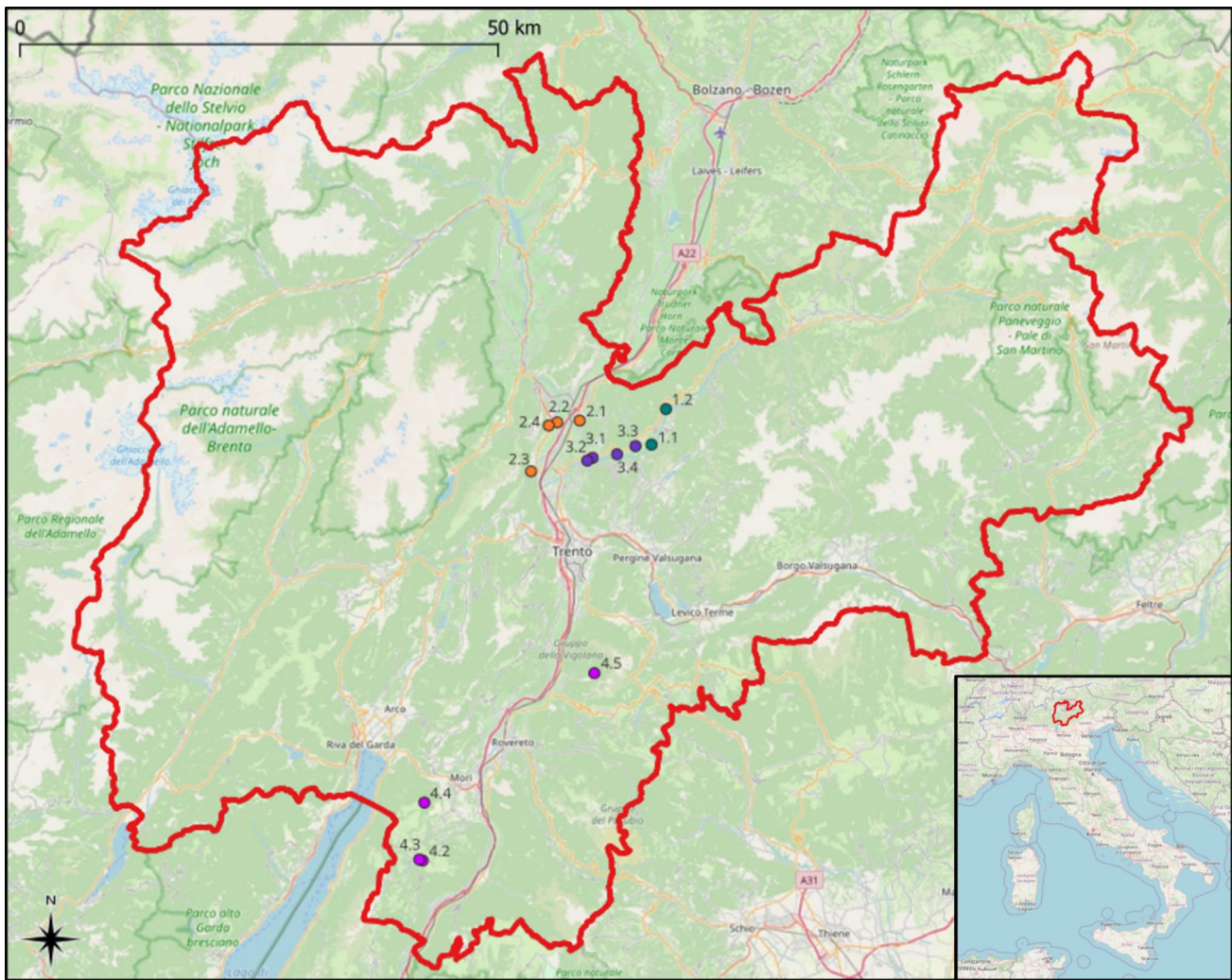


Fig. 1 Map of the study area (Province of Trento) with its location on the Italian peninsula (lower right panel). Sample sites are represented by closed circles, colored according to habitat type (FLU: dark green, VBE: orange, AGR: purple, PAS: pink) and labeled according to site

code numbers (see Table 1). Scale bar and compass are also shown on the left. This image was created with QGIS.org (2024) QGIS Geographic Information System. Open Source Geospatial Foundation Project <http://qgis.org>

Bioinformatics and Statistical Analysis

Raw reads processing and taxonomic assignment were performed with the R package *DADA2* V1.30.0 [43, 44] using default parameters, except for the following modifications. For the 16S bioinformatic workflow, at the *filterAndTrim* step the option *trimLeft* was set as *c(17,21)* and *truncLen* as *c(280,250)*. The database used for taxonomic assignment was *silva_nr99_v138.1* [45]. The initial raw reads were 21,407,900 and after bioinformatic processing 10,318,149. For the ITS1 bioinformatic workflow, primer trimming was performed with *cutadapt* by assessing all possible primer orientations to ensure complete primer removal. Taxonomic assignment was performed with the UNITE database version 9.0 (UNITE general FASTA release for eukaryotes, released on 18.07.2023; [46]). In this case, the initial dataset

consisted of 8,481,680 raw reads, and after bioinformatic processing 3,683,426. At the end of both workflows, all data were converted to a *phyloseq* object [47] for further analysis. The prevalence-based method of the R package *decontam* (v1.20.0; [48]) was applied to detect and filter contaminating ASVs detected from negative controls. The number of retrieved ASVs at the end of all bioinformatic steps was 50,984 for bacteria and 7343 for fungi.

The following statistical analyses were performed with the R packages *microeco* v1.3.0 [49] and *vegan* v2.6.4 [50]. Fungal ASVs matching *Bd* (ncbi ACC ID: NR_119535.1) were identified using the UNITE-based taxonomic classification and confirmed by BLASTn [51]. Association between habitat subtypes and *Bd* detection was tested with Chi-squared test with the R package *stats* v4.3.2. As the presence of bacteria with antifungal activity has been shown

to inhibit *Bd* infection, to detect bacteria with putative anti-fungal activity in our samples, the bacterial ASVs found in our samples were mapped onto the sequence database by Woodhams et al. [14]. Only sequences spanning the entire V3-V4 16S rRNA gene region annotated as *Bd*-inhibitory bacteria were included in the analysis (649 sequences). Skin and water ASVs were matched to the corresponding *Bd*-inhibitory sequences trimmed to the V3-V4 16S rRNA gene portion using the closed reference clustering strategy implemented in MICCA [52]; sequence identity threshold: Correlation between the presence and relative abundance of each bacterial ASV (including ASVs with putative *Bd*-inhibitory activity) and the following three variables: detection of *Bd*, relative abundance of *Bd*-related ASVs, and the number of *Bd*-related ASVs, was tested using Spearman's rank correlation coefficient with the R package `corrplot` V0.95.

The number of ASVs shared between water and skin microbiota within and between habitat subtypes was estimated using the *microeco* function "trans_venn" after merging all libraries according to the variable (e.g., habitat subtype, habitat, and sample type). ASVs associated with significant differences in their abundance across habitat subtypes were identified with a differential abundance testing approach carried out using non-normalized data with the `ALDEx2_t` function implemented in *microeco* [53]. *p* values were corrected for false discovery rate using the Benjamini–Hochberg correction (significance cutoff of FDR corrected *p* values: 0.05). Heatmaps were generated using the web tool ClustVis [54] considering bacterial and fungal ASV counts normalized using relative proportions (total sum scaling, TSS) multiplied by median sequencing depth.

To estimate standard alpha diversity indices (Faith's Phylogenetic Diversity, hereafter PD, Chao1, and Shannon diversity), bacterial and fungal libraries were rarefied to 21,598 and 2200 reads per sample, respectively, thereby losing one bacterial and six fungal libraries. Differences in alpha diversity estimates across habitat subtypes were tested using nonparametric Dunn's Kruskal–Wallis Multiple Comparisons test with Holm *p*-value adjustment and using a significance level of 0.05.

We used linear mixed models (LMM) with the R package *lme4* [55] to investigate the impact of water temperature, dissolved oxygen, pH, and water bacterial and fungal diversity (Shannon) on alpha-diversity estimates of skin bacteria and skin fungi (as separate response variables). For these analyses, alpha diversity indices were transformed using the R package *bestNormalize*, applying Box–Cox transformation to bacterial and fungal Chao1 estimates, arcsin transformation to bacterial PD, and Ordered Quantile (ORQ) normalization (`orderNorm`) to bacterial Shannon. Fungal Shannon estimates did not significantly deviate from a normal distribution (Shapiro–Wilk test, $W = 0.98658$, *p* value = 0.2199)

and were not transformed. All explanatory variables were scaled using the function `scale` in R package *dplyr* [56].

To estimate beta-diversity indices (Bray–Curtis and Jaccard), bacterial and fungal samples were first normalized using a scaling with ranked subsampling (SRS, [57]). Unconstrained ordination based on Bray–Curtis and Jaccard dissimilarity estimates was performed using non-metric multidimensional scaling (NMDS). Beta-diversity estimates were then compared across habitat subtypes using ANOSIM and permutational MANOVA (PERMANOVA) statistical tests with 999 permutations. Multivariate homogeneity of group dispersions, a necessary condition to apply the PERMANOVA test, was tested using the *vegan* `betadisper` function [58] in the R package *microeco*. The correlation between both bacterial and fungal beta-diversity indices with pH, temperature, and dissolved oxygen as well as water bacterial and fungal alpha diversity (Shannon) were tested with a Mantel test in R using the Spearman correlation coefficient. The same statistical test was also used to check for correlation between bacterial and fungal beta diversity estimates.

Results

Overall, this study of *B. variegata* skin microbiota detected 88 microbial Phyla, including 74 bacteria and 14 fungi. The most common bacterial Phyla in both skin and water communities were Proteobacteria, Bacteroidota and Verrucomicrobiota, followed by Cyanobacteria, Actinobacteriota, and Patescibacteria (see Fig. 2). The most common fungal Phylum in skin swabs was Ascomycota, followed by Basidiomycota (Fig. 2). Of note, while taxonomic classification at Phylum level in skin samples exceeded 75% of sequence reads in most cases (Fig. 2), the majority (median 82.5%; range: 14.4–99.4%) of fungal reads in water samples could not be classified even at this high taxonomic level (Fig. 2).

Despite sampled toads lacking noticeable symptoms of pathogen infection, the taxonomic classification of fungi in skin samples led to the identification of 26 ASVs matching *Bd* (Table S1), which were not detected in our water samples. The 18 *Bd*-positive skin samples (12% of collected swabs) spanned all four habitat subtypes (8/14 sites): five were from FLU (AMB_1_1, AMB_1_2), three from PAS (AMB_4_3, AMB_4_5), three VBE (AMB_2_2), and seven from AGR (AMB_3_2, AMB_3_3, AMB_3_4). Overall, *Bd* detection was not statistically associated with any of the studied habitat subtypes (Pearson's chi-squared test, *p* value = 0.08987).

Using publicly available databases [14], we identified 63 bacterial taxa (ASVs) with putative anti-*Bd* activity (Table S2). Almost a third of these were classified as

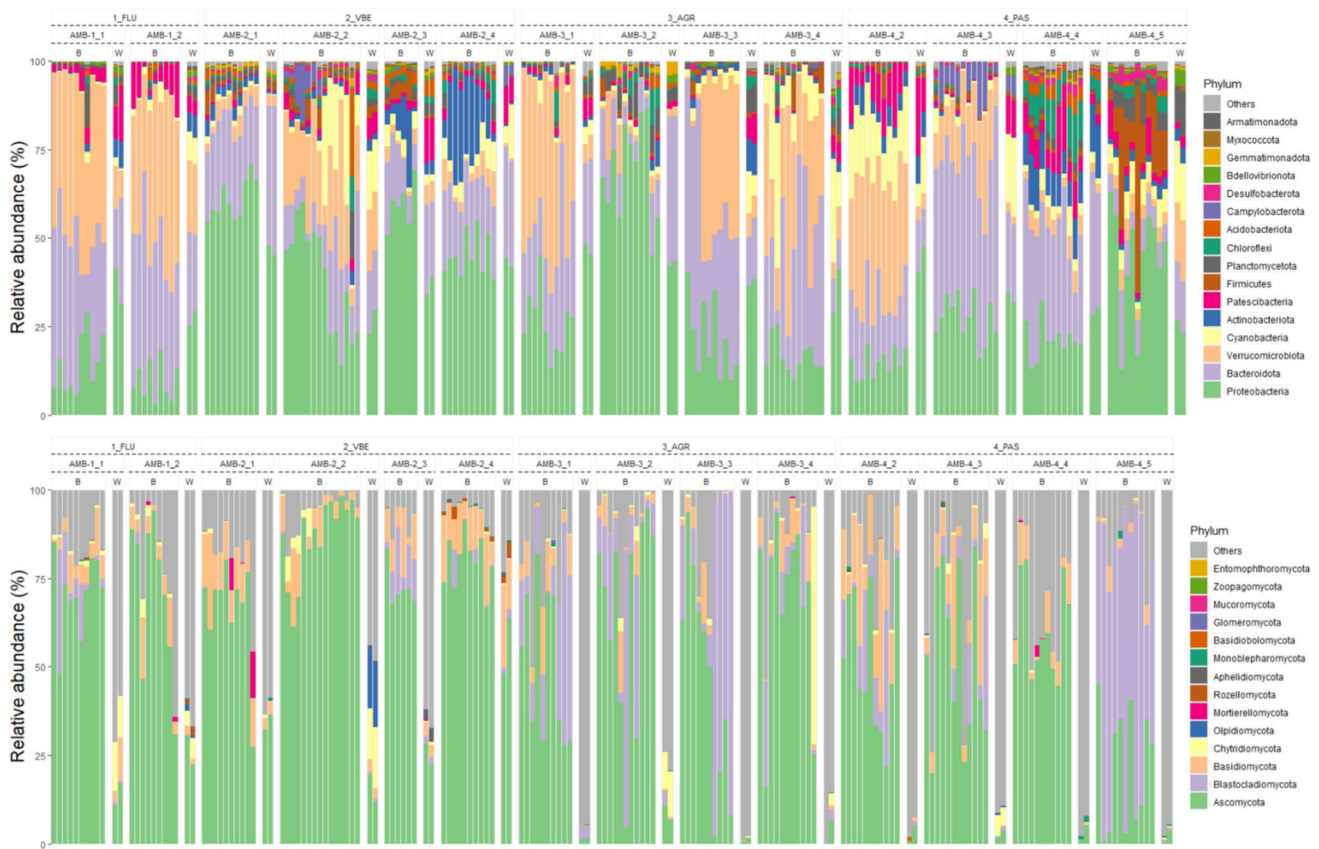


Fig. 2 Relative abundance of Phyla of *Bombina variegata* skin (B) and water (W) microbiota. Upper panel: bacteria; lower panel: fungi. Upper bar refers to habitat subtypes (1_FLU, fluvial sites, 2_VBE, valley bottom ecotones, 3_AGR, agricultural water tanks, 4_PAS, pasture ponds); codes in the second bar to sampled sites (AMB-1_1

etc.); third bar indicates the origin of the sample: *B. variegata* skin (B) or water (W). This plot was created with the R package *ggplot2* and formatted using *GIMP* v2.10.18 (The GIMP Development Team (2019). GIMP. Retrieved from <https://www.gimp.org>)

Pseudomonas spp. (18 ASVs). Other genera represented by three or more distinct ASVs were *Acinetobacter* (5 ASVs), *Stenotrophomonas* (5 ASVs), *Chryseobacterium* (3 ASVs), *Flavobacterium* (3 ASVs), and members of the family Aeromonadaceae (3 ASVs). These taxa were found in both water and skin samples, from all four habitats. However, the relative abundance or presence of putative *Bd*-inhibitory ASVs did not show any correlation with the detection of *Bd* in skin samples, the number of *Bd* related ASVs, or the cumulative relative abundance of *Bd* related ASV (Spearman correlation coefficient ranging from -0.15 to $+0.15$ considering both relative abundance and presence of putative *Bd*-inhibitory ASVs; Figure S1, Table S2). Moreover, extending this analysis to the entire set of bacterial ASVs identified in this study did not highlight significant correlations between abundance of skin bacterial ASVs and detection of *Bd* in skin samples, the number of *Bd* related ASVs or the cumulative relative abundance of *Bd* related ASVs (Spearman correlation coefficient ranging from -0.21 to $+0.43$).

The comparison of bacterial and fungal ASVs found in skin and water in each habitat subtype revealed that the microbiota of water and skin communities were quite distinct (Figure S2A, B), with a relatively low number of shared bacterial and fungal ASVs. Nonetheless, these accounted for more than half of generated reads in each habitat (range: 55.5 to 80.9% for bacteria; 60.2 to 76.8% for fungi). Skin microbial communities, defined as ASVs detected only from skin samples in each habitat, included 32,022 bacterial and 2774 fungal ASVs (Figure S2C, D) representing 69.57% and 68.60% of identified bacterial and fungal ASVs. Of these, only about 1% (e.g., 302 bacterial ASVs and 39 fungal ASVs) were detected in at least one sample from each habitat subtype (Figure S2C, D). However, again, these few ASVs shared between skin and water accounted for approximately 63% and 31% of total reads sequenced from bacterial and fungal skin communities, respectively. The comparison of water and skin microbiota using bacterial ASVs collapsed at Phylum level revealed that while a significant fraction of identified Phyla was only detected in skin or water samples

(with the number of non-shared Phyla ranging from 12/66 for AGR to 15/50 for FLU), these private Phyla accounted for less than 1% of sequence reads, highlighting that shared Phyla were by far the most abundant.

Overall, 17 bacterial ASVs showed significant differences in abundance in skin microbial communities among habitat subtypes (Fig. 3, Table S3). These differentially abundant taxa suggested a clear separation between VBE and the other habitats (Fig. 3A) as well as the separation of FLU from AGR and PAS (Fig. 3B). Of note, ten, five, and three taxa found in FLU skin samples showed significant differences in abundance compared to the same taxa in VBE, PAS and AGR samples, respectively (Table S3, Fig. 3A). Conversely, only three fungal ASVs classified as *Aureobasidium* (Ascomycota), *Claviceps* (Ascomycota), and *Catenaria* (Blastocladiomycota) showed significant differences across habitat subtypes (Table S3).

The alpha diversity estimates calculated from *B. variegata* skin swabs varied between habitat subtypes (Fig. 4), with the skin microbiota of individuals from FLU and AGR showing significantly lower bacterial PD and Chao1 richness estimates than skin samples collected from the other two habitats (VBE, PAS). Shannon diversity estimates for bacterial communities were significantly different across all four habitats, with VBE having the highest diversity followed

by PAS, AGR, and FLU (Dunn's Kruskal–Wallis Multiple Comparisons, $p < 0.05$; Fig. 4, Table S4). In addition, for each of the three diversity indices, significant differences were detected between skin bacterial communities hosted by animals from different sampling sites within habitat subtypes AGR and PAS, but not FLU and VBE (Dunn's Kruskal–Wallis Multiple Comparisons, $p < 0.05$; Figure S3).

Fungal richness and diversity, considerably lower than those observed for bacteria, also varied significantly between habitat subtypes, with skin mycobiota of VBE toads characterized by higher Chao1 and Shannon diversity estimates than those sampled in PAS and AGR (Dunn's Kruskal–Wallis Multiple Comparisons, $p < 0.05$; Fig. 4, Table S4). Moreover, similarly to what was observed for bacterial communities, Chao1 and Shannon diversity estimates of skin fungal communities varied between sampling sites within each habitat subtype except PAS, for which we found no significant difference for Shannon diversity estimates (Dunn's Kruskal–Wallis Multiple Comparisons, $p < 0.05$; Figure S3).

Linear mixed models indicated that observed variation in skin bacterial richness (Chao1) and PD across habitats and sampling sites was negatively associated with pH and positively associated with measured levels of dissolved oxygen (Table S5). Notably, the same associations were not detected when bacterial alpha diversity was estimated

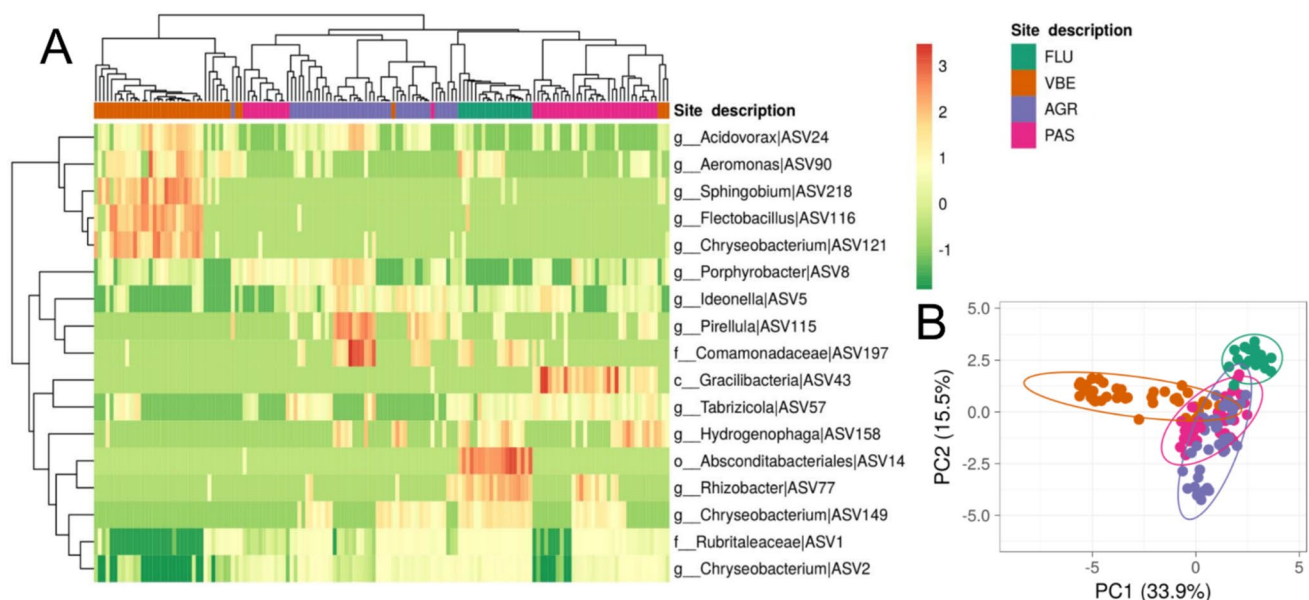


Fig. 3 Heatmap (A) and PCA (B) of the 17 bacterial taxa of *B. variegata* skin microbiota with significant differences in abundance across four habitat subtypes (FLU, VBE, AGR, and PAS). To generate both A and B, original values of ASV abundance were $\ln(x+1)$ -transformed. In (A), rows were centered, unit variance scaling was applied to rows, and both rows and columns were clustered using correlation distance and average linkage (74 rows, 147 columns). Abundance is represented with colors from green (rank 1; least abundant) to red (rank 3; most abundant). For the PCA, unit variance scaling

was applied to rows, and singular value decomposition (SVD) with imputation was used to calculate principal components. X- and Y-axes show principal components 1 and 2 that explain 33.9% and 15.5% of the total variance, respectively. Prediction ellipses are such that with probability 0.95, a new observation from the same group will fall inside the ellipse. This plot was created with ClustVis and formatted using GIMP v2.10.18 (The GIMP Development Team (2019). GIMP. Retrieved from <https://www.gimp.org>)

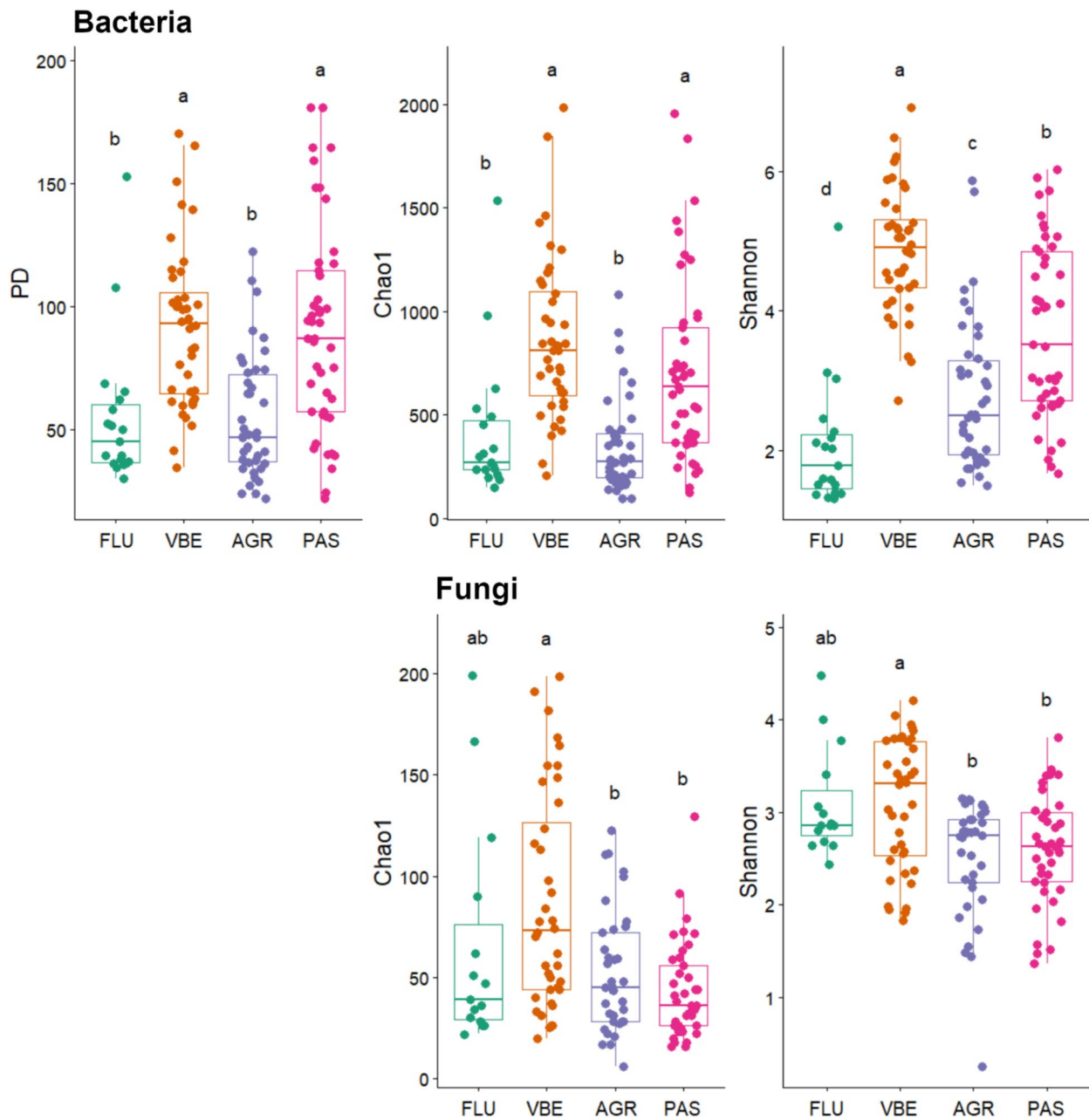


Fig. 4 Alpha diversity estimates (PD: Faith's phylogenetic diversity, Chao1 and Shannon) of *B. variegata* skin microbiota across four habitat subtypes (FLU, VBE, AGR, and PAS) in the province of Trento, Italy. Upper panel: bacteria; lower panel: fungi. This plot was created

with the R package *ggplot2* and formatted using *GIMP* v2.10.18 (The GIMP Development Team (2019). GIMP. Retrieved from <https://www.gimp.org>)

as Shannon diversity, suggesting that rare or low abundance taxa might be vulnerable to these abiotic changes. Interestingly, variation in the same abiotic factors was not associated with changes in fungal diversity (both indices, Table S5). Instead, the observed Shannon diversity, but not richness (Chao1), of fungal communities was negatively correlated with variation in temperature across the investigated habitats and sampling sites.

The composition of skin and water microbiota, as measured by Bray–Curtis and Jaccard dissimilarity indices, varied significantly between habitat subtypes (Fig. 5: skin samples; Figure S4: skin and water samples). For both indices, the four subtypes were characterized by differences in dispersion of bacterial skin microbiota (BETADIPER, Bray–Curtis F : 40.50, p value ≤ 0.001 ; Jaccard F : 6.2695, p value ≤ 0.001), as highlighted by the close clustering of FLU bacterial

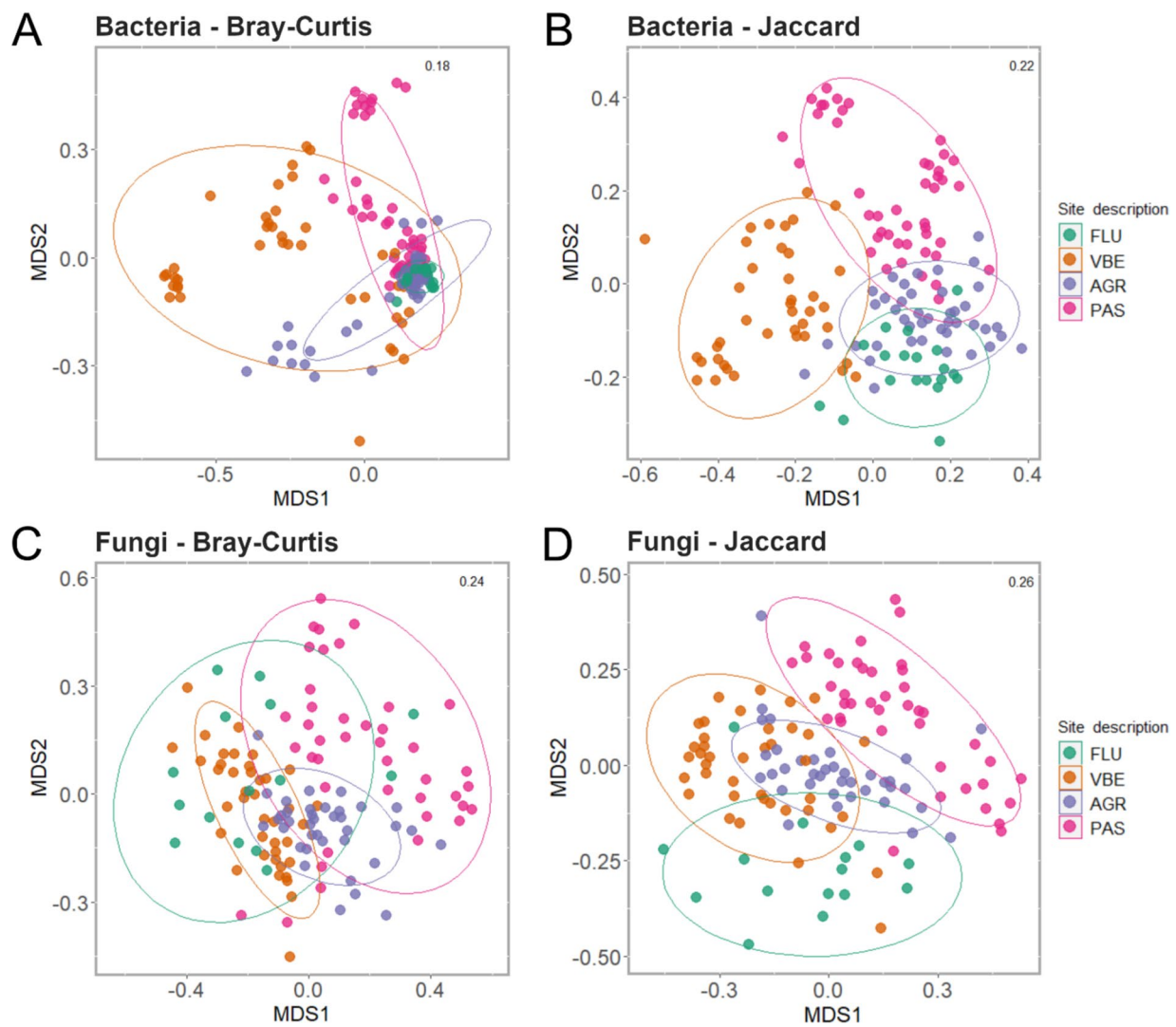


Fig. 5 Beta diversity estimates (left: Bray–Curtis; right: Jaccard) of *B. variegata* skin microbiota across the four habitat subtypes (FLU, VBE, AGR, and PAS). Upper panel: bacteria; lower panel: fungi.

This plot was created with the R package *ggplot2* and formatted using *GIMP* v2.10.18 (The GIMP Development Team (2019). GIMP. Retrieved from <https://www.gimp.org>)

communities compared to the widespread distribution of VBE and PAS communities (Fig. 5A, B). However, pairwise dissimilarity estimates and clustering of samples according to both indices showed that skin bacterial communities of animals collected from the same habitat were more similar to one another than to those of other habitats (ANOSIM, Bray–Curtis R : 0.248, p value ≤ 0.001 ; Jaccard R : 0.601, p value ≤ 0.001). Consistent with this result, both dissimilarity estimates highlighted significant differences across skin bacterial communities hosted by animals collected from different habitat subtypes (PERMANOVA, Bray–Curtis R^2 : 0.183, F : 20.03, p value ≤ 0.001 ; Jaccard R^2 : 0.075, F : 4.50, p value ≤ 0.001 ; Tables S6, S7 and Fig. 5). In particular, our results show that VBE is clearly differentiated from the other subtypes, especially in terms of relative abundances of skin

bacterial taxa, followed closely by FLU (Table S7). In terms of presence/absence data of bacterial taxa, however, all subtypes appear to be more equally distinct (Table S7). Nevertheless, a significant fraction of variation in dissimilarity estimates was found to be associated with differences across sampling sites within each habitat subtype (PERMANOVA, Bray–Curtis R^2 : 0.414, F : 13.56, p value ≤ 0.001 ; Jaccard R^2 : 0.190, F : 3.40, p value ≤ 0.001 ; Fig. 5, Table S6), with skin bacterial communities found in VBE and PAS being characterized by the highest differentiation across sampling sites according to both indices (Fig. 5).

The ordination of skin fungal communities based on Bray–Curtis and Jaccard dissimilarity indices grouped the skin samples into distinct clusters corresponding to the four habitat subtypes. Again, separation between subtypes was

evident when ecological dissimilarity between communities was estimated with the Bray–Curtis (ANOSIM R : 0.3643, p value ≤ 0.001 ; PERMANOVA R^2 : 0.126, F : 3.450, p value ≤ 0.001 ; Table S6) and Jaccard indices (ANOSIM R : 0.5204, p value ≤ 0.001 ; PERMANOVA R^2 : 0.074, F : 3.450, p value ≤ 0.001 Fig. 5C, D and Figure S4C, D; Tables S6, S7). Furthermore, as reported above for bacterial communities, a large proportion of variation in diversity estimates was associated with differences across sampling sites considered in each habitat subtype (PERMANOVA, Bray–Curtis R^2 : 0.231, F : 4.27, p value ≤ 0.001 ; Jaccard R^2 : 0.154, F : 2.37, p value ≤ 0.001 ; Table S6, Fig. 5C, D). Notably, both beta diversity indices show an association between skin communities and water communities from the same site for both bacteria and fungi (PERMANOVA p value ≤ 0.001 for both sample type and interaction term between site code: sample type; Figure S4; Table S6), although overlap in community composition is higher when dissimilarity was estimated with Jaccard index (Figure S4; Table S6).

Mantel tests highlighted that water temperature had the strongest correlation with observed differences in bacterial (Spearman correlation: Bray–Curtis $R=0.53$, $p=0.002$ and Jaccard $R=0.43$, $p=0.001$) and fungal diversity estimates (Spearman correlation: Bray–Curtis $R=0.31$, $p=0.002$ and Jaccard $R=0.30$, $p=0.002$) (see Table S8). Additionally, we found evidence of correlation between dissolved oxygen and skin bacterial diversity (Spearman correlation: Bray–Curtis $R=0.25$, $p=0.002$ and Jaccard $R=0.20$, $p=0.001$) as well as skin fungal diversity (Spearman correlation: Bray–Curtis $R=0.20$, $p=0.002$ and Jaccard $R=0.27$, $p=0.002$), while other parameters showed much lower correlation coefficients (Table S8). It is interesting to note, however, that the composition of bacterial and fungal diversity appears to be poorly correlated, but that the correlation is stronger when the comparison between pairs of samples is made using presence/absence data. (Mantel test, Bray–Curtis R : 0.08093, p value: 0.037; Jaccard R : 0.1583; p value: $1e-04$).

Discussion

Skin microbial communities may be fundamental for amphibians to adapt to challenging environments such as polluted or highly seasonal water bodies. This study aimed, for the first time, to characterize the skin microbiota of the yellow-bellied toad, a locally threatened amphibian, across a variety of habitat subtypes. Despite the potential conservation relevance and the number of papers on bacterial skin communities, this is one of the first amphibian microbiota studies focusing on both bacterial and fungal skin diversity. Overall, although core taxa for both skin bacterial and fungal communities were similar to those previously reported separately, we found an unexpected association between the

diversity indices of both microbiota components. These indices varied significantly across habitat subtypes, and there were strong associations between skin bacteria and fungi and those of the water from each habitat subtype, as well as with water parameters.

The main bacterial Phyla characterizing the skin microbiota of the yellow-bellied toad, Proteobacteria, Bacteroidota, and Verrucomicrobiota also dominate the skin microbial community of other amphibian species (e.g., *B. orientalis* [59], *Thoropa taophora* [60], and five species of the *Pseudoeurycea* genus [61]). Since several Proteobacteria species are known to produce anti-*Bd* metabolites, the abundance of this taxon on *B. variegata* skin may explain the apparently low susceptibility of this toad to *Bd*, as has been suggested for another species from warmer climate (*T. taophora* from Brazil [60]). The high prevalence of Ascomycota and Basidiomycota is also in agreement with previous studies on amphibian skin mycobiota [9, 60, 61]. Interestingly, most fungal taxa found in our water samples were classified as “unknown” (i.e., not present in the standard database for metataxonomy of fungi, UNITE). The lack of data on wetland fungal diversity has in fact been noted by several authors [62]. Considering the clear association of wetland fungal communities with that of the yellow-bellied toad, and by extension, to all amphibian species with life stages dependent on water bodies, we suggest further research focused specifically on water mycobiota, in order to deepen our understanding of its impact on freshwater organisms.

Worryingly, for the first time in the Province of Trento, we detected ASVs from amphibian skin swabs showing 100% identity with the nucleotide sequences of *Bd* reference specimens (NCBI ACC: NR_119535.1). Several individuals in each of the four habitat subtypes were positive for *Bd*, but there was no clear pattern of association of *Bd* presence with subtype. All individuals with these *Bd* ASVs appeared asymptomatic. The apparent lack of symptoms of the sampled individuals to *Bd* infection could be due to the presence of 63 putative *Bd*-inhibitory bacterial taxa. This number is also likely to be an underestimate of the richness of inhibitory taxa since most of the reference sequences in Woodhams et al. [14] were obtained from other continents (e.g., North America, New Zealand) and only 89 were from Europe (Switzerland). On the other hand, no clear correlation was found between presence or relative abundance of *Bd*-related ASVs and bacterial ASVs, including those matching putative *Bd*-inhibitory bacteria. Nevertheless, we should note that the low number of *Bd*-infected samples, their distribution among different habitat subtypes, and the high bacterial diversity among habitat subtypes made the identification of clear correlations between the incidence of this pathogenic fungus and changes in specific bacterial taxa particularly challenging. Moreover, *Bd* virulence has been shown to depend on strain (e.g., [63]), and recent

reports suggest that European strains may be less virulent than those of other continents, and possibly even protect amphibians from colonization with more lethal ones [64]. This could explain why wild populations of yellow-bellied toad appear to be highly tolerant to this pathogen. However, as landscape context has been identified as a major aspect shaping *Bd* distribution, changes in land use and water quality could alter this precarious coexistence in the near future [65]. Therefore, our result, if confirmed, is highly relevant for amphibian conservation in northern Italy [32].

In agreement with previous studies (e.g., two salamander genera, *Ensatina* and *Batrachoseps* [6]), our skin samples showed a high proportion of unique ASVs (not found in water samples; Figure S2), despite the toads' skin being in direct contact with the water microbiota, suggesting that skin microbiota composition is highly selected by the host. In fact, only a low number of ASVs were shared between skin and water, although shared taxa were present with a high relative abundance in both sample types (Figure S2). Similar numbers have been reported for other amphibian species (e.g., *Eleutherodactylus johnstonei* [21]). Nonetheless, the overall *B. variegata* skin microbial community appears to be largely composed of the relatively few, but highly abundant, ASVs that are also found in their environment, thus supporting the influence of habitat subtype on skin microbiota and highlighting once more the importance of habitat conservation for these sensitive species. As both bacterial and fungal communities have shown a strong association with water microbiota communities, we suggest that conservation strategies for *B. variegata*, and amphibians more in general, include environmental microbiota and the impact of anthropogenic activities in these environments.

Our skin samples highlighted several differentially abundant bacterial taxa between habitats, including Gammaproteobacteria as well as Bacteroidia, Alphaproteobacteria, Gracilibacteria, Planctomycetes, and Verrucomicrobiae. These taxa present different characteristics in terms of habitat preferences and metabolism; therefore, further investigation will be needed to understand the significance of this composition. In addition, both alpha and beta-diversity estimates for skin bacteria and fungi were significantly different for individuals living in distinct habitat subtypes, confirming recently published results that found, for other amphibian species, variations in skin community composition according to habitat type (e.g., *T. thaofora* [60], *Anaxyrus boreas* [17], and eight other species [51]). In particular, individuals of VBE sites showed a higher diversity for both microbial communities and both indices. Several authors have shown the association between a rich and diverse microbiota and healthier individuals [e.g., 66]. Therefore, we argue that *B. variegata* populations living in ephemeral ponds along the bottom of Adige Valley are potentially healthier and more resistant to *Bd*

[67], although a more comprehensive study with a higher number of samples will be needed to test this hypothesis, as well as to provide relevant insights on *Bd* presence in the area (i.e., which strain is present).

Alpha and beta diversity indices for bacterial and fungal communities also showed strong correlations with the environmental parameters considered here. For example, bacterial richness (alpha diversity) was positively correlated with dissolved oxygen in the water and negatively correlated with water pH. While the negative impact of water pH confirms results from other studies [25, 68], the positive effect of dissolved oxygen was less expected (i.e., [68] found a moderate negative correlation for this variable) but could help explain the lower diversity found in individuals sampled in ponds subjected to eutrophication (i.e., pasture ponds, PAS). Both bacterial and fungal beta diversity were positively correlated with water temperature and, although less strongly, with dissolved oxygen. Again, the positive effect of temperature we found is in agreement with previous studies, although available only for bacterial communities [25, 68], while the positive influence of dissolved oxygen represents a novel finding. Therefore, for future *B. variegata* populations conservation planning, our results support the need to preserve and possibly increase freshwater sites where waters are moderately warm and oxygenated, and eutrophication is limited.

Since this is a relatively unexplored field, our results are timely as they show more specific associations between water parameters and skin microbiota composition and richness. García-Sánchez and colleagues [61] reported that fungal skin community composition of amphibians appears to be even more strongly influenced by climatic conditions than that of bacterial communities, which they found more dependent on host phylogenetic relatedness. Overall, the association of skin and environment microbial communities seems so strong that they could be considered meta-communities, as recently suggested by Leonhardt et al. [21]. For this reason, we support a holistic approach to the study of amphibian microbiota, that includes both bacterial and fungal communities of both host and environment.

Supplementary Information The online version contains supplementary material available at <https://doi.org/10.1007/s00248-025-02489-1>.

Acknowledgements The authors would like to thank the Fondazione E. Mach for access to facilities, and Sonia Endrizzi, Matteo Trenti and Luca Roner for sharing their knowledge of the field sites to include in the study and for fieldwork support.

Author Contribution This study was conceived and designed by L.Z., M.G., P.P., G.B., and H.C.H. Funding was acquired by M.G., P.P., G.B., and H.C.H. The samples were collected by L.Z., M.G., S.C., G.G., and I.L.P., while laboratory analyses were performed by L.Z. and M.G. Bioinformatic and statistical analyses were carried out by L.Z. and G.G. The first draft of the manuscript was written by L.Z., G.G. and H.C.H., and all authors read, commented and approved the final manuscript.

Funding Open access funding provided by Fondazione Edmund Mach - Istituto Agrario di San Michele all'Adige within the CRUI-CARE Agreement. L.Z. was supported by a PhD scholarship cofunded by the University of Ferrara and MUSE–Science Museum Trento. H.C.H. and M.G. were partially funded by the Ledro Alps and Judicaria UNESCO Biosphere Reserve (Acqua e Vita project) and HCH at the Fondazione E. Mach (Project BIOALPEC) by the National Biodiversity Future Centre (NBFC) Project (code CN_00000033, Concession Decree No. 1034 of 17 June 2022 adopted by the Italian Ministry of University and Research, CUPD43C22001280006), funded under the National Recovery and Resilience Plan (NRRP), Mission 4 Component 2 Investment 1.4-Call for tender No. 3138 of 16 December 2021, rectified by Decree n.3175 of 18 December 2021 of Italian Ministry of University and Research funded by the European Union–NextGenerationEU financed by the Ministero dell'Università e della Ricerca, PNRR, Missione 4 Componente 2, “Dalla ricerca all’impresa,” Investimento 1.4, Progetto CN00000033 (BIOALPEC).

Data Availability The datasets generated and analyzed during the current study have been submitted to the EMBL Database (Project Accession: PRJEB81466), and will be made publicly available upon acceptance of the article.

Declarations

Ethics Approval Animal handling procedures for research purposes were authorized by the Italian Ministry of the Environment according to DPR 357/97 on April 7th, 2021 (Prot. ISPRA 16759 of 6/4/21; Authorization no. 0035657 of 7/4/21 by the Ministry of the Environment).

Competing Interests The authors declare no competing interests.

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